

## Supplementary material

### Plant species supplementary information

**Table S1.** Summary of traits of exotic (bold type) and native seeds. Mean and standard deviation of traits measured in 50 seeds per species. In the case of seed number per fruit, we counted seeds in 50 fruits per species.

Species	Width (mm)	Length (mm)	Area (mm <sup>2</sup> )	Seeds per fruit	Seed mass (g)	Coat hardness*	Shape
<i>A. microphylla</i>	1.74±0.31	1.84±0.25	2.57±0.67	1.62±0.89	0.002±0.001	1	Spherical
<i>R. idaeus</i>	<b>1.96±0.56</b>	<b>2.36±0.53</b>	<b>3.49±0.78</b>	<b>27.6±5.93</b>	<b>0.003±0.001</b>	<b>2</b>	<b>Oval</b>
<i>A. chilensis</i>	3.25 ± 0.26	2.59±0.38	6.75±1.17	3.64±1	0.012±0.002	1	Spherical
<i>M. boaria</i>	3.99±0.56	2.02±0.49	6.82±1.12	2±0	0.012±0.004	3	Oval
<i>B. darwinii</i>	4.11±0.43	2.09±0.27	6.76±1.2	4.68±1.25	0.005±0.001	2	Oval
<i>S. patagonicus</i>	4.57±0.41	4.41±0.50	15.95±2.36	1±0	0.033±0.005	2	Spherical
<i>P. cerasus</i>	<b>10.11±0.55</b>	<b>8.37±0.47</b>	<b>66.61±6.47</b>	<b>1±0</b>	<b>0.457±0.051</b>	<b>3</b>	<b>Spherical</b>

\*Seeds coat hardness categories: 1: could be penetrated easy by a fingernail; 2: could not be penetrated by a fingernail; 3: difficult to penetrate with a knife (Gong et al. 2015).

### Estimates of relative abundance of seeds at landscape scale

To estimate the relative abundance of seeds we combined data of the relative abundance of fruiting plants of each species and their canopy size (data from transects), their fruit production and phenology (data from focal individuals) and the number of seeds becoming available for granivores (data from tagged branches in focal individuals). All this information was combined into a model that simulated the number of individuals present at the landscape scale and their fruit production (according to allometric functions). Then, for each week we simulated the number of sound fruits (according to each species' phenology) and the number of seeds available on the ground (considering fruit consumption by dispersers and fallen due to fruit ripening).

In December 2019, when all the focal species were in bloom, we established 5 transects 100 m long x 2 m width (1000 m<sup>2</sup>) randomly located throughout the landscape (10 ha, mean distance across transects 200 m). In each one, we registered adult individuals of fleshy-fruited species, and their canopy projection (diameter, m). With this information we obtained a distribution of canopy sizes for each species. In dioecious species (*S. patagonicus*, *A. chilensis*, and *A. microphylla*) we identified the sex of flowers to consider sex-ratios when estimating the number of trees producing fruits. In addition, we selected 10-15 focal tree individuals from each species and measured the canopy diameter. For each focal tree, at the beginning of the fruit bearing season, we counted the number of fruits present in the canopy, randomly marked 10 branches and counted the number of fruits. On a weekly basis and for each tagged individual of each species, we

monitored the number of fruits present (in branches and in the canopy) and the proportion of fruits missing in the branches (with respect to previous week). Missing fruits were mostly due to fall of ripe fruits or dispersal by birds (clean vs pulp remaining in the pedicel). To estimate the number of seeds per fruit, for each focal species, we collected a total of 50 fruits across 5 to 10 individuals and counted the number of seeds.

Following Carlo and Morales (2016) we estimated the number of fruits per species by means of bootstrapping techniques. In each iteration, we first sampled a number of individuals equal to those present in 10 ha (according to the relative abundances observed in transects) and assigned them a canopy size (sampled from the Gaussian kernel density estimated from the data of the landscape transects). When necessary, we corrected this number by the sex ratio of species. For instance, if there were 20 individuals of the b-th species in 10 ha we sampled 20 values of canopy size from the Gaussian kernel density estimated from the observed canopy sizes for the i-th species. If the species had a sex-ratio of 0.5 we sampled 10 values of canopy size. Then, for each sampled individual we simulated fruit production as a function of their canopy size (based on allometric functions fitted to observed data in focal trees). Allometric functions were obtained by regressing the number of fruits produced by focal individuals against their canopy size. We used four candidate functions of fruit production-canopy relationships: (i) intercept, no relationship; (ii) lineal, (iii) saturating and (iv) exponential. For each species, we selected the more suited allometric function based on AIC values. On a weekly basis, we simulated fruit ripening according to species phenology ( $F_{jkw} = Crop_{jk} \times P_{kw}$ ). The number of sound fruits of the j-th individual, corresponding to species k, during the w-th week ( $F_{jkw}$ ) was obtained by multiplying its total fruit production ( $Crop_{jk}$  by  $P_{kw}$ ), which represents species-specific fruiting phenology.  $P_{kw}$  is the ratio between the average number of ripe fruits produced by individuals of the k-th species in the w-th week and the average maximum fruits they produce. From week to week, we simulated the number of seeds becoming available for granivores due to frugivory (and subsequent defecation) or natural fall. The number of fruits becoming available in the j-th focal tree during the w-th week ( $f_{jkw}$ ) was equal to the number of ripe fruits in the previous week ( $F_{jw-1}$ ) multiplied by the average proportion of fruits missing from the previous week in branches of tagged individuals of the k-th species ( $\overline{M_{kw}}$ ).  $f_{jkw} = F_{jw-1} \times \overline{M_{kw}}$ . The number of fruits was converted into seeds by multiplying these values by the average number of seeds per fruit of each species. Finally, for

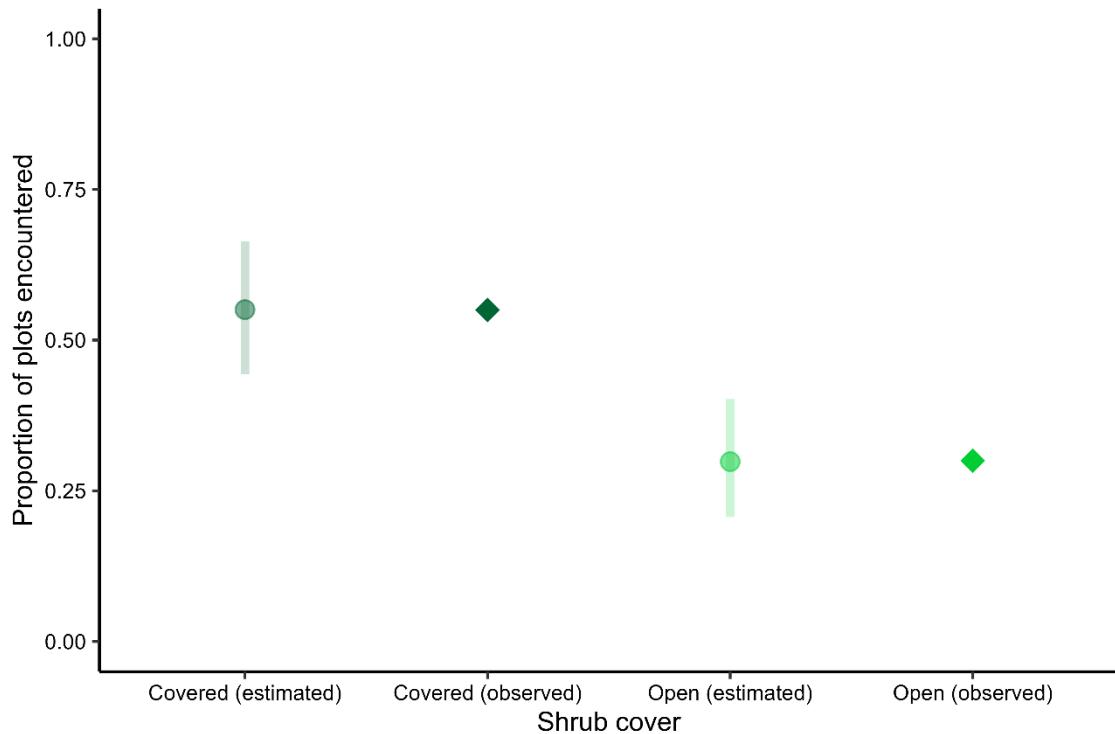
each species, we summed seeds of all individuals and divided this value by the sum across species to calculate the relative abundance. Simulations were repeated 2000 times and averaged per species and week.

### Model specifications

#### 3.1. Effect of shrub cover on plot encounter probability

$$\begin{aligned}
 Y_{is} &\sim \text{Bernoulli}(p_i) \\
 \text{logit}(p_{is}) &= \beta_{0s} + \beta_1 * O_i \\
 \beta_{0s} &\sim \text{Normal}(\mu_0, \sigma_0) \\
 \mu_0 &\sim \text{Uniform}(-\infty, \infty) \\
 \sigma_0 &\sim \text{Student t}(0, 2.5, 3)
 \end{aligned}$$

The probability that at least one seed from the  $i$ -th plot and  $s$ -th site, was removed during the first 24 h of seeds offering, was modeled as a Bernoulli process that depended on the probability of encountering the plot  $p_{is}$ .  $p_{is}$  was a function of the microhabitat where the plot was located (0 = covered; 1 = open). The  $s$ -th site in which the plot was situated was introduced as a random effect on the intercept ( $\beta_{0s}$ ). We used weakly informative priors for the intercept and sigma (random effect)—student-t distribution with mean = 0, df = 3 and sd = 2.5. For the mean of the random effect ( $\mu_0$ ) we used a uniform distribution  $(-\infty, +\infty)$ . We ran 3 chains over 10000 iterations, leaving 500 for warm-up, and a thinning of 10. After checking for convergence ( $\text{Rhat} < 1.01$ ) and effective sample sizes ( $\text{n.eff} > 2600$ ), we evaluated model fit by means of posterior predictive checks (Fig. S2). We fitted the model using the brms 2.17.0 R package (Bürkner 2017).



**Figure S1.** Posterior predictive check of the model for plot encounter probability (Table 2, main text). Rhombus depict observed values and dots mean predictions and 90% credible intervals across 1,000 posterior samples.

### 3.2. Effects of shrub cover on seeds removal rates

$$Y_{isw} \sim \text{Binomial}(p_{isw}, N)$$

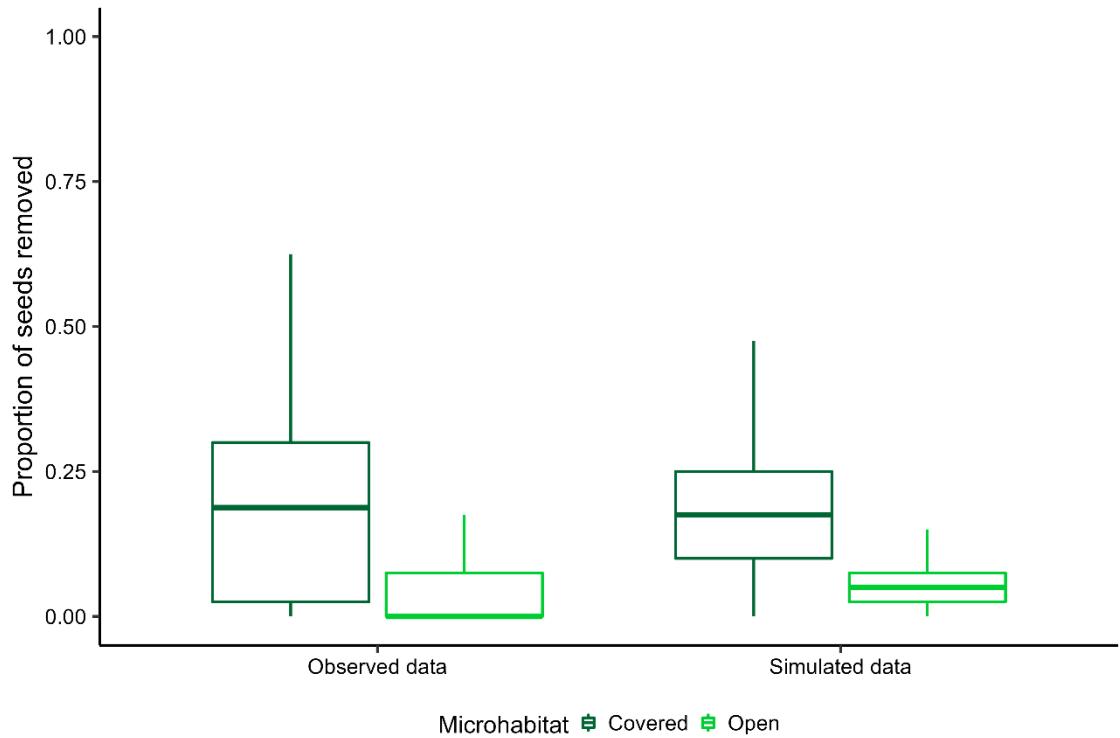
$$\text{logit}(p_{isw}) = \beta_{0_s} + \beta_1 * O_i$$

$$\beta_{0_s} \sim \text{Normal}(\mu_0, \sigma_0)$$

$$\mu_0 \sim \text{Uniform}(-\infty, \infty)$$

$$\sigma_0 \sim \text{Student } t(0, 2.5, 3)$$

The number of seeds removed in the  $i$ -th plot,  $s$ -th site, during the  $w$ -th week was modeled as a binomial process with probability of seed removal  $p_{isw}$  and number of seeds offered ( $N = 40$ ). The probability of seed removal depended on the microhabitat where the plot was located (0/1, covered by shrubs or open areas). The  $s$ -th site in which the plot was situated was introduced as a random effect on the intercept ( $\beta_{0_s}$ ). We used weakly informative priors for the intercept and sigma (random effect)—t student distribution with mean = 0, df = 3, sd = 2.5. For the effect term and the mean of the random effect ( $\mu_0$ ) we used a flat prior, a uniform distribution ( $-\infty, +\infty$ ). We ran 3 chains over 10000 iterations with 1000 of warm-up and a thinning of 10. After checking for convergence ( $\text{Rhat} < 1.01$ ) and effective sample sizes ( $\text{n.eff} > 2500$ ), we evaluated model fit by means of posterior predictive checks (Fig. S2). Model was fitted using Brms 2.17.0 program (Bürkner 2017).



**Figure S2.** Posterior predictive check of the seed removal model (Table 2, main text). Box plots represent observed and simulated values (across 1000 repetitions) of the proportion of seed removed after 72 h of seed offering. Dark green boxplots represent plots located under shrub cover microhabitats, and light green ones, open microhabitats.

### 3.3. Factors driving foraging choices by granivores

$$y_{kiw} \sim \text{WalleniusHipergeometric}(p_{kiw}, T_{iw}, N_{kiw})$$

$$p_{kiw} = \frac{\exp(\eta_{kiw}) * I_{kw}}{\sum_{j=1}^7 \exp(\eta_{jiw}) * I_{jw}}$$

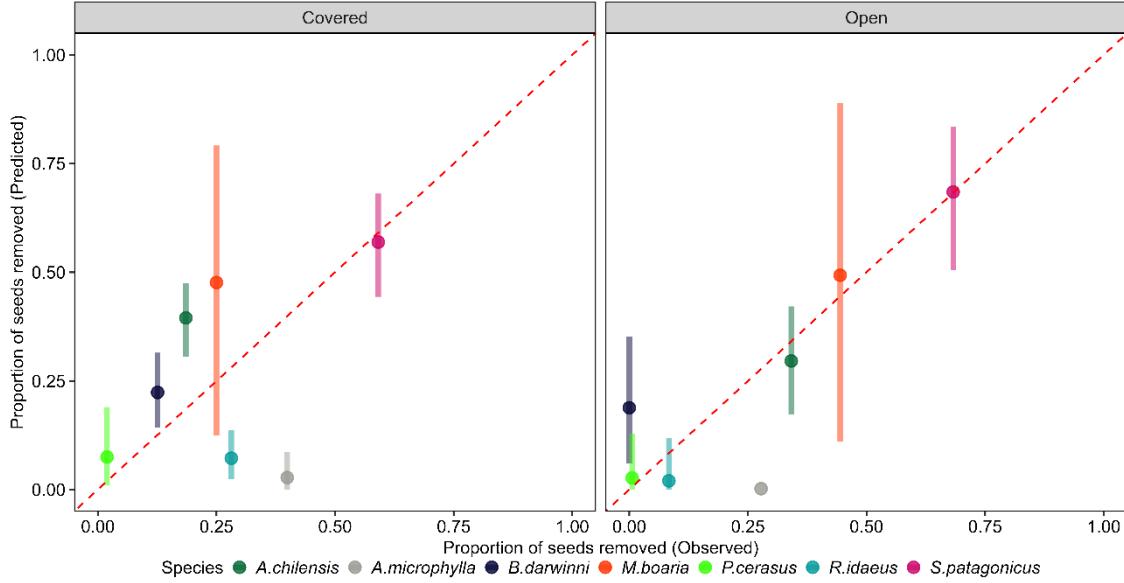
$$I_{jw} \begin{cases} 0 & \text{if the } I - \text{th species is absent} \\ 1 & \text{if the } I - \text{th species is present} \end{cases} \quad I_{kw} \begin{cases} 0 & \text{if the } I - \text{th species is absent} \\ 1 & \text{if the } I - \text{th species is present} \end{cases}$$

$$\eta_{kiw} = \beta_1 S_k + \beta_2 S_k^2 + \beta_3 A_{kw} + \beta_4 S_k * O_i + \beta_5 S_k^2 * O_i + \beta_6 A_{kw} * O_i$$

$$\beta_n \sim \text{Normal}(0,1) \text{ for } n \in \{1, 2, \dots, 6\}$$

The number of seeds removed of the  $k$ -th species in the  $i$ -th plot during the  $w$ -th week followed a multivariate Wallenius' Hipergeometric distribution with a species-specific probability of removal  $p_{kiw}$ , a total number of removed seeds  $T_{iw}$  and number of seeds offered by species  $N_{kiw}$ . Species-specific probability of seed selection depended on their size, following a quadratic response ( $\beta_1 S_k + \beta_2 S_k^2$ ), their estimated species-specific relative abundance per week ( $\beta_3 A_{kw}$ , log-transformed) and the interaction of both terms with shrub cover. Covariates were standardized (mean=0,  $sd = 1$ ) before model fit. We used

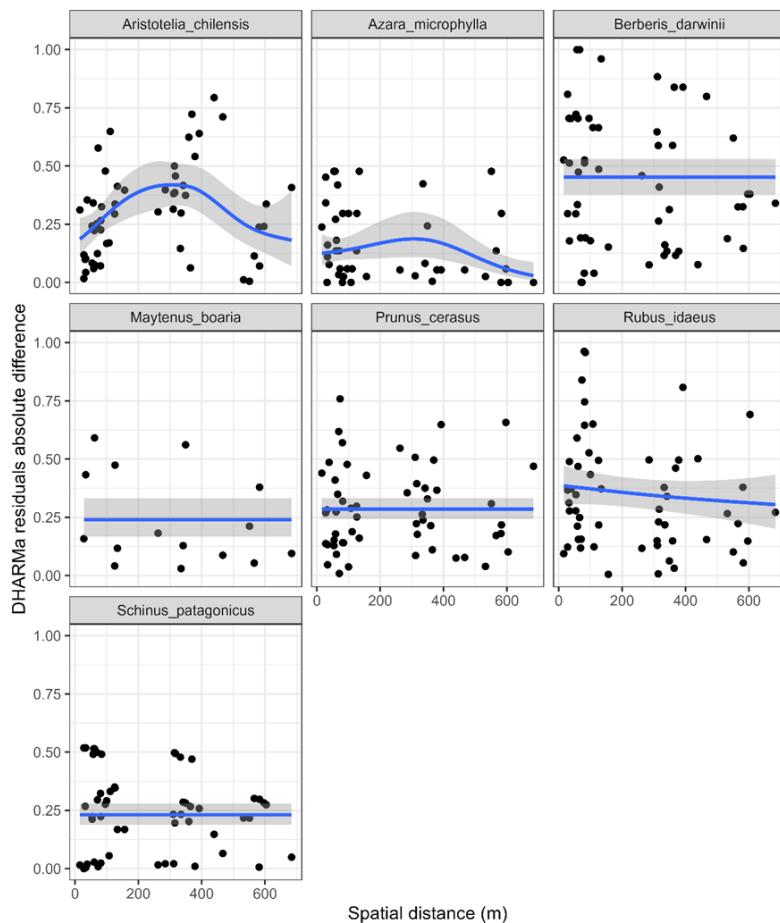
weakly informative priors for all parameters – normal distribution with mean = 0, sd = 1. We run 3 chains over 200000 iterations with a thinning of 20. After checking for convergence (Rhat <1.01) and effective sample sizes (n.eff >9000), we evaluated model fit by means of posterior predictive checks (Fig. S3). Model was fitted using the R package BiasedUrn (Fog, 2022) for computing the likelihood function, and sampling the posterior distribution with random walk Metropolis algorithm, as implemented in the MCMC R package (Geyer, 2022).



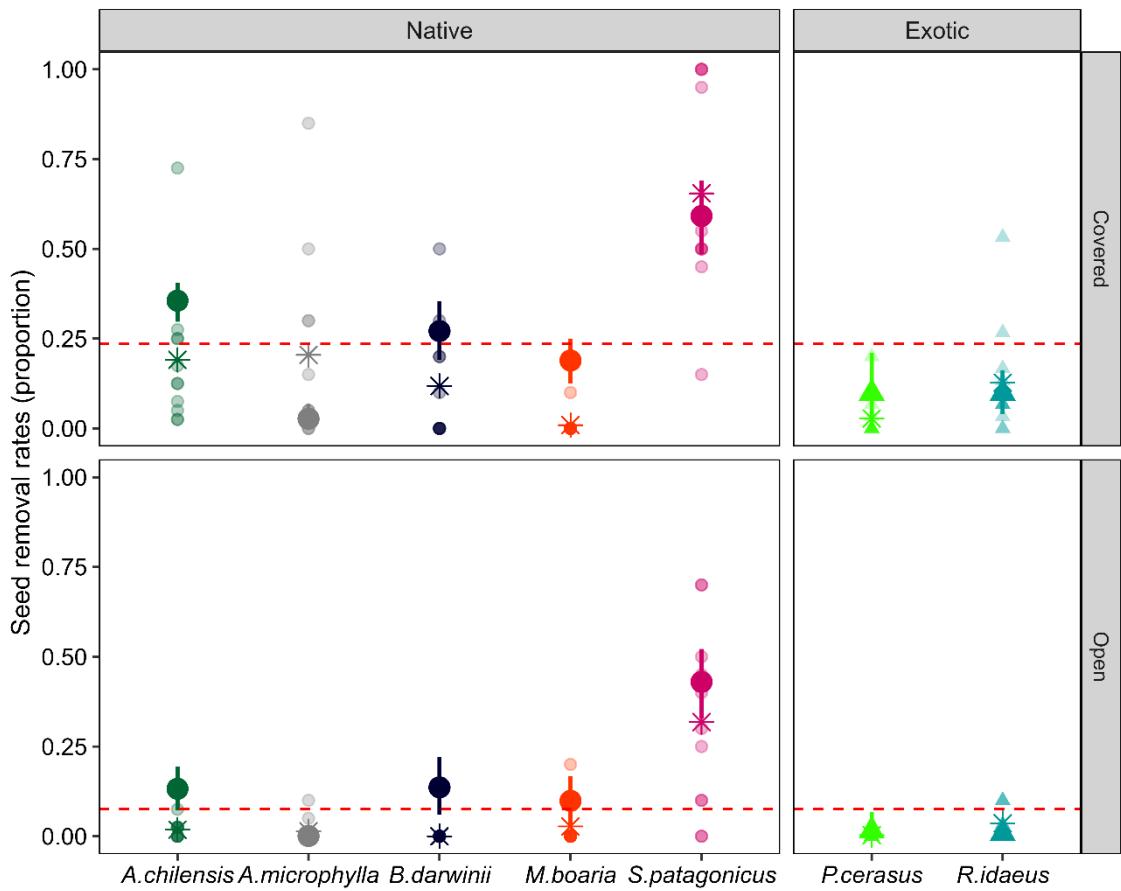
**Figure S3:** Posterior predictive check of the seed selection model (Table 3, main text). Dots represent the mean predicted value of the seed removal rates after 72 h of offering (across 10000 posterior samples), and lines, the 90 % credible interval. The red line shows the 1:1 relationship.

**Table S2.** Results of the Moran test for spatial autocorrelation calculated on the quantile residuals ("DHARMA" package in R). In all cases the p-value > 0.05 suggest the absence of spatial autocorrelation between sites.

Model	Moran's index	p-value
<b>Probability of encounter</b>	0.21	0.14
<b>Removal rates</b>	-0.18	0.24
<b>Seed selection</b>	-0.07	0.82



**Figure S4.** Spatial correlation analysis of the residuals of the hypergeometric regression adjusted to data (Table 3). Points are the absolute pairwise difference between the DHARMA residuals averaged by site as a function of the pairwise spatial distance between sites for each species. The line and ribbon represent a beta generalized additive model fitted to the data, with its 95 % confidence interval. The lack of a positive trend suggests the absence of spatial correlation.



**Figure S5.** Total seed removal rates by species throughout the 4 weeks of the cafeteria experiment. Native and exotic species are represented with circles and triangles, respectively. Upper panels represent removal rates in plots located under shrub cover, lower panels represent patterns in open microhabitats. Light dots depict raw data and asterisk the mean values across plots throughout the season. The red line represents the mean removal rates of native species. Dark dots and lines represent the mean and 90% credible interval predicted by the full Hypergeometric model (Table 3).